

A COMPARISON OF RANGELAND MONITORING TECHNIQUES FOR RELATIVE
SPECIES ABUNDANCE IN NORTHERN MIXED GRASS PRAIRIE

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A Comparison of Rangeland Monitoring Techniques for Relative Species Abundance in
Northern Mixed Grass Prairie

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ABSTRACT

Clipping by species is one of the most accurate methods available for determining species composition. However, cost and time constraints often make clipping by species impractical on any large scale. Our objective was to determine whether either of two less labor intensive methods (line point intercept, quadrat frequency frame) could provide suitable alternatives to clipping. Data was collected as part of a rangeland monitoring project on the Dakota Prairie Grasslands. Each method was used to inventory grassland plant species within the same plots, which allowed us to analyze the results for each method side by side. Our findings indicate that for relative rankings of species abundance, both line point intercept and quadrat frequency frames produce a similar result as clipping and could be used interchangeably. We suggest using either line point intercept or quadrat frequency frames to produce such a list because of the reduced time inputs involved.

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LIST OF ABBREVIATIONS

AVHRR.....	Advanced Very High Resolution Radiometer
BLM.....	Bureau of Land Management
DPG.....	Dakota Prairie Grasslands
LCTA.....	Land Condition-Trend Analysis
MLRA.....	Major Land Resource Area
NDVI.....	Normalized Difference Vegetation Index
NMS.....	Nonmetric Multidimensional Scaling
NRCS.....	Natural Resource Conservation Service
NRI.....	Natural Resource Inventory
USFS.....	United States Forest Service
VOR.....	Visual Obstruction Readings

INTRODUCTION

Rangelands and grasslands cover approximately 50% of the earth's terrestrial surface (Williams et al. 1968). These areas provide vital forage for livestock as well as crucial wildlife habitat. Less obvious but equally important functions of rangeland include carbon sequestration, water cycling, and erosion control (Havstad et al. 2007). Land managers must monitor rangeland vegetation to determine whether their management decisions are producing desired results; however, detailed monitoring of these huge tracts of land is both expensive and time consuming.

Some methods, such as photo-points or visual cover estimation, take little time or expertise to perform, but give only very basic information. Other methods, like above-ground biomass clipping, can give very detailed and accurate information about a rangeland plant community, but are time consuming and therefore expensive (Bonham 1989). In some cases, detailed information may be necessary. In many other situations a relative species abundance ranking will be adequate. This ranking can give managers a fairly comprehensive species list, identify dominant species, and recognize invasive species. They can also be produced with less time consuming methods than clipping. Our objective for this study was to determine whether these less time consuming methods could provide a suitable alternative to clipping in large scale monitoring projects. If this proves true, land managers, land owners, and researchers would have the option of increasing frequency of monitoring (versus clipping), increasing amount of land monitored, or simply reducing the time and monetary input required for monitoring.

We propose that techniques requiring a moderate time input, such as quadrat frequency frames and line point intercept methods, provide the best balance between detailed information and time input when monitoring large areas. To test this hypothesis we compared the relative plant community composition hierarchies produced by two moderate input methods (line point intercept

and quadrat frequency) to those produced by clipping by species in two common ecological sites in western North Dakota.

LITERATURE REVIEW

Quadrat Frames

Quadrat frames are a commonly used tool in modern plant community composition studies. They are used in many ecosystems including wetlands (e.g., Lopez and Fennessy 2002), forests (e.g., Archaux et al. 2007), and rangelands (e.g., Mosley et al. 1989). Quadrat frames have been used since at least 1912 (Priestly 1913), with further use in rangelands as early as 1914 (Ramsey 1916).

The quadrat frame has been used in many different ways. Sampson (1917) listed three ways to use the quadrat: mapping the size and location of each plant within the frame, listing each species within the frame, or removing all plants from the frame and recording the repopulation. Weaver and Clements (1929) described use of quadrats as grids and also as a random sample. They also discussed varying the size and number of quadrats based on type of vegetation studied. For example, larger quadrats would be used in a sagebrush community than in a mixed-grass prairie. Also, more samples should be taken in a diverse, variable community than are necessary in a relatively homogeneous community.

Wiegert (1962) discussed the harmful edge effects seen when using a quadrat that is too small. He also created a method for determining the optimum quadrat size and shape using calculated variance at different quadrat sizes. Despain et al. (1991) recommended using a square quadrat of 40-50 cm per size in most locations. They also advocated use of nested quadrats of varying size to monitor species of varying sizes and frequencies. These researchers also stated that “the best sampling precision is reached for a particular species when it is present in 40% to 60% of the quadrats sampled. This will provide the most sensitivity to changes in frequency.” West (1985) warned that managers should not attempt to calculate percent composition from frequency data, and Hironaka (1985) stated that other methods should be used in conjunction with quadrat frequency if time and money are available.

Daubenmire (1959) described a quadrat method for determining canopy cover. He used 6 canopy cover classes to quantify plant communities. However, Floyd and Anderson (1987) and Kennedy and Addison (1987) recommended that quadrats not be used for cover estimation. Cain (1943) used the quadrat frame as a method for determining plant communities based on frequency, and Curtis and McIntosh (1950) described various methods for analyzing quadrat data, including frequency, density, abundance, constancy, and presence. Quadrat frames are useful in recording rare species, as the area covered by the method is quite large. Time inputs vary based on use, with canopy cover mapping being quite time consuming while frequency determination requires less time. Frequency monitoring does have the problem of weighting a single species within the quadrat the same as a species that appears several times. Greig-Smith (1983) determined that the relationship from frequency to density is only consistent when plants are evenly distributed. Hironaka (1985) found that plant frequency monitoring was faster and cheaper than any other method if your goal is detecting changes in community composition. Recently, researchers have continued to use the quadrat frame for vegetation monitoring and are also working to compensate for sampling bias in the method (Clarke et al. 2011). Heywood and DeBacker (2007) found that a plot size producing an average of about 50% frequency yields nearly maximum statistical power, and that revisiting the same sites over time increased statistical power, which agrees with the findings of Despain et al. (1991). DeBacker et al. (2011) used differential plot sizes ranging from 0.01 m² to 10 m² to reach an optimal plot size for species growing in a tallgrass prairie system. They found that in order to reach an optimum frequency for detecting changes, plot sizes needed to vary between common and less common species of concern. They recommend determining the optimal plot size for each species of concern and using plots of that size to monitor changes over time.

Harvest Method: (Clipping Above-ground Biomass)

Clipping has been used as a monitoring technique since at least the late 1800's. Fream (1888, 1890) removed turf and soil samples from the field and grew them in a laboratory to maximum production, before clipping the above ground mass and weighing it without drying. Stapleton (1913) used a clipping technique to measure production, also weighing green vegetation. Roberts (1933) dried the green biomass before weighing in order to obtain a more accurate production value. Clipping, particularly by separating each species in the sample, has several advantages. It is very accurate and removes much of the observer bias that is seen in other techniques. The method also gives an accurate portrayal of the biomass production of the area, which is especially useful in setting stocking rates for grazing animals. In recent years, clipping by species has been used in rangeland management for similarity indices. These indices compare the production of each species at a given site with the probable historical production of native species.

An issue with clipping by species is the amount of time required to sort through all of the plant species in a given clipping. Bonham (1989) noted that biomass estimation by harvesting is "time consuming and expensive." This is especially true when working in a diverse plant community. Because of this problem, several researchers have tried to correlate clipping data with less time consuming methods, especially visual obstruction readings (VOR). Benkobi et al. (2000) and Uresk and Mergen (2012) both found correlations between above ground production and VOR, but their correlation coefficients varied by ecosystem. Ahmed et al. (1983) and Brummer et al. (1994) tested the efficacy of different quadrat shapes, but no shape was found to be superior in accuracy. Clark et al. (2008) compared the point intercept method with clipping in sagebrush steppe ecosystems and found point intercept as accurate as clipping but less time consuming. Thoma et al. (2002) used Advanced Very High Resolution Radiometer (AVHRR) Imagery to produce a Normalized Difference Vegetation Index (NDVI), which they then compared to clipping data, again in an

attempt to reduce the time and energy expended in monitoring. They found that AVHRR had the potential to predict forage values at a regional level, but had no ability to determine species composition. Epstein et al. (2012) compared clipping data to satellite derived Normalized Difference Vegetation Indices (NDVI) to attempt to improve the NDVI system in tundra ecosystems. They found that they needed extensive ground level harvesting to strengthen the model.

Another issue with biomass clipping is the time required to transport samples to the laboratory for drying and the time spent drying. This transportation and weighing process creates more chances for data to be mislabeled or misplaced. The equipment needed (e.g., clipping hoops or frames, clippers, bags, labels) also adds significant weight to the researcher's pack when working in remote locations.

Point Intercept (Multi and Single Point Methods)

Everson and Clarke (1987) stated that “when a sampling quadrat is reduced to a dimensionless point, frequency becomes an absolute measure of cover.” The point method of vegetation monitoring was developed in New Zealand by Levy and Madden (1933). The first frame consisted of a wooden horizontal bar supported by strips of steel. Ten steel needles were inserted through the wooden bar at two inch intervals. Any plant contacted by a pin at any height was recorded. Often, the height at which the contact was made was recorded. Modifications to the frame to add a brake system (Heady and Rader 1958) and a hinged leg for easier storage and angled reading (Smith 1959) helped make the point frame easier to use. Wilson (1960) found that angling points at 32.5 degrees produced the least variation from foliage angle. Nerney (1960) added a bicycle tire to the frame, but that improvement has not been as popular. To measure both canopy and basal cover, a single pin is often substituted for the pin frame. In this variation of the method, all plant species that contact the pin are recorded, as well as the ground cover that the tip of the pin touches.

Early verification of the point intercept method's validity was done by Goodall (1952), Whitman and Siggeirson (1954), and Heady (1957). Since then the single point intercept method has been used extensively in the Natural Resource Inventory (NRI) performed by the Natural Resource Conservation Service (NRCS) to measure both canopy and basal cover. Several other federal agencies use the method including the Bureau of Land Management (BLM) and the United States Forest Service (USFS). The ten point intercept method has become one of the most common methods for determining ground cover.

Ecological Sites

An ecological site is defined as “a distinctive kind of land with specific physical characteristics that differs from other kinds of land in its ability to produce a distinctive kind and amount of vegetation” (Sedivec and Printz 2012, USDA-NRCS 2006). Our study area comprised many different ecological sites, but the two most common ecological sites were the loamy and the thin loamy ecological sites. Therefore, we chose to focus on these sites in our analysis.

The loamy ecological site is the most common ecological site in North Dakota. These sites are on uplands with a surface soil layer that forms less than a 2-inch ribbon of silt loam or loam. Subsoil layers form a less than 2-inch ribbon of silt loam to clay loam. The upper part of the subsoil is none to slightly effervescent (Sedivec and Printz 2012). The plant community is comprised of about 85% grasses and grass-like plants, 10% forbs, and 5% shrubs. The plant community is dominated by western wheatgrass (*Pascopyrum smithii*) and green needlegrass (*Nassella viridula*), with needle-and-thread (*Hesperostipa comata*), blue grama (*Bouteloua gracilis*), porcupine grass (*Hesperostipa spartina*), and sedges (*Carex spp.*) common as well. Common forbs include American vetch (*Vicia americana*), green sagewort (*Artemisia dracunculoides*), silverleaf scurfpea (*Pediomelum argophyllum*) and Missouri goldenrod (*Solidago missouriensis*). Western snowberry (*Symphoricarpos occidentalis*) is the

principal shrub and occurs in patchy mosaic. In other areas, silver sagebrush is the dominant shrub and occurs more evenly dispersed across the site (USDA-NRCS 2003). The combination of gentle slopes and optimal available water content make these sites one of the most productive ecological sites in MLRA (Major Land Resource Area) 54 and 58C, with annual production averaging about 2400 kg/ha.

The thin loamy ecological site was the second most common site in our study area. In a thin loamy ecological site the surface layer forms a less than 2-inch ribbon of silt loam or loam. The subsoil layer forms a ribbon of silt loam to clay loam less than 2-inch in length. These soils have none to strong effervescence in the surface layer and strong to violent effervescence in the subsoil. These sites occur on ridges and knolls (Sedivec and Printz 2012).

The plant community is comprised of about 85% grasses or grass-like plants, 10% forbs, and 5% shrubs. The major grasses include the needle-and-thread, green needlegrass, western wheatgrass, little bluestem (*Schizachyrium scoparium*), and sideoats grama (*Bouteloua curtipendula*). Other grasses occurring on the site include blue grama, plains muhly (*Muhlenbergia cuspidata*), and red threeawn (*Aristida purpurea*). The increased slope, and consequently increased runoff, of the thin loamy site leads to less production, averaging about 1500 kg/ha in MLRA 54 and 58C.

STUDY AREA

Our study took place on the Dakota Prairie National Grasslands of western North Dakota, which are managed by the United States Department of Agriculture - United States Forest Service (USFS). These National Grasslands cover over 500,000 hectares and are split into four ranger districts: Medora Ranger District, McKenzie Ranger District, Sheyenne Ranger District, and Grand River Ranger District. Our data were collected in the McKenzie and Medora districts by researchers based in three research regions: McKenzie County, Billings County, and Slope County.

Historically, the grasslands of this region were grazed by bison, elk, and other native herbivores (Hanson 1984). They were also subjected to disturbance from fire (Wells 1970). In the past 150 years, cattle have been the dominant large herbivore on these grasslands. Cattle grazing has been a major component of the USFS multiple-use strategy since the agency was formed in 1891. Fire has also been used but is not as common and often met with resistance. In our study area, and many other National Grasslands, grazing has been managed cooperatively by the USFS and local grazing associations. These associations are made up of landowners and lessees (permittees) who graze federal lands, and were created to prevent smaller ranchers from being pressured to sell by large corporations or the federal government. This relationship has not been without conflict, but research-based, cooperative decisions can help reduce controversy, especially when stocking rates are involved. The grazing associations within our study area included the Little Missouri, McKenzie, Medora, and Grand River Grazing Associations.

Our study locations were located within MLRAs 54 and 58C. MLRA 54 is described as rolling soft shale plain (USDA-NRCS 2006). The soil parent material of the region is soft, calcareous shales, siltstones, and sandstones. The dominant soil orders of this MLRA are Mollisols and Entisols, and the northern and eastern parts of the area have been modified by glaciation. The average annual precipitation of MLRA 54 is 355-455 mm and average annual temperature 3-8

degrees C. Nearly all of this MLRA is farmed or ranched, and about half of the area supports native grasses and shrubs that are grazed. The natural prairie vegetation expected would be northern mixed-grass prairie, with green ash (*Fraxinus pennsylvanica*), chokecherry (*Prunus virginiana*), and buffaloberry (*Shepherdia argentea*) occurring in draws and ravines. The vegetation of this region supports many wildlife species including whitetail deer, mule deer, pronghorn antelope, prairie dogs, and many bird species.

MLRA 58C is described as the northeastern part of the northern rolling high plains. This area consists of rolling hills and badlands. The soil parent material includes marine sediments, shale, siltstone, and sandstone. Mollisols are present, but Entisols and Inceptisols are common due to a widespread occurrence of steep slopes and erosion caused by these slopes. The average annual precipitation of this area is 355-430 mm and average annual temperature 5-7 degrees C. MLRA 58C land use is dominated by ranching and recreation. Only 5% of the MLRA is cropland (USDA-NRCS 2006). Northern mixed-grass prairie occurs in most of the area, with barren badland outcrops being fairly common. This MLRA also supports many species of wildlife.

Ecological Sites

We limited our analysis to the two most common ecological sites to ensure an adequate sample size in each. Our monitoring plots were selected to assess the status of the two most common ecological sites in each allotment, which provided the most data on loamy and thin loamy ecological sites.

The loamy ecological site is the most common ecological site in North Dakota. These sites are on uplands with silt loam to clay loam textured soils. These are typically well developed soils with few remaining carbonates at the surface (Sedivec and Printz 2012). The historic climax plant community of this ecological site includes green needlegrass, western wheatgrass, blue grama, and several native forb species. The combination of gentle slopes and optimal available water content

make these sites one of the most productive ecological site in MLRA 54 and 58C, with annual production averaging about 2400 kg/ha.

The thin loamy ecological site was the second most common site in our study area. Thin loamy ecological sites display loam to clay loam soils that are less well developed than those found on loamy sites. These soils often contain significant amounts of carbonates and occur on ridges and knolls (Sedivec and Printz 2012). The historic climax plant community of thin loamy sites includes little bluestem, western wheatgrass, sideoats grama, sedges, and several species of more drought tolerant forbs and shrubs. The increased slope, and consequently increased runoff, of the thin loamy site leads to less production, averaging about 1500 kg/ha in these MLRAs.

Precipitation

Growing season precipitation in the southern portion of our study area (Figure 1) was slightly above average. In the northern portion (Figure 2) growing season precipitation was somewhat below average. We feel that this amount of variation from the mean is normal and has had no effect on our comparison of methods.

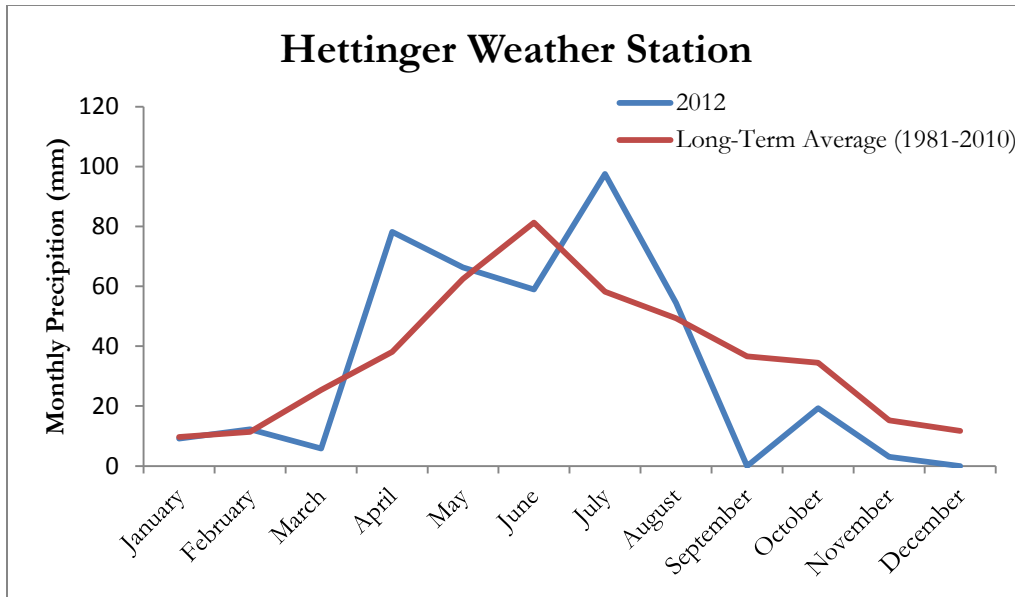


Figure 1. Monthly precipitation (mm) at Sidney, MT weather station for 2012 and long-term average (USDC Commerce 2012a).

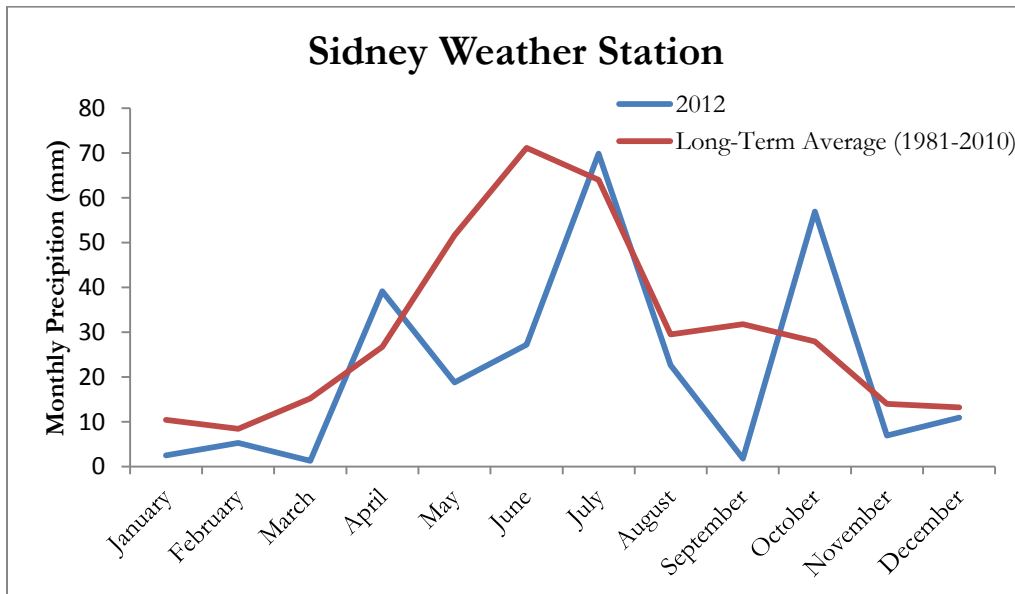


Figure 2. Monthly precipitation (mm) at Hettinger, ND weather station for 2012 and long-term average (USDC Commerce 2012b).

METHODS

Plot Location

Monitoring plots for a large scale study on the Dakota Prairie Grasslands (DPG) were located using a systematic approach followed by random selection. Each allotment or grazing unit was stratified by ecological site using the USDA-NRCS soil layers for McKenzie, Billings and Slope counties in North Dakota, and Perkins county South Dakota. Once the ecological site layer was created, the two dominate ecological sites, in terms of acres, were selected for data collection. Three or more plots were selected randomly for each ecological site within each grazing allotment. Two hundred meter buffer exclusion was created from the roads, and 100 m buffer exclusion from the water tanks, fences, and pipelines.

From these plots listed in the previous paragraph, the loamy and thin loamy ecological sites were chosen for this study. The loamy and thin loamy ecological sites were two of the most common ecological sites found on the DPG. In all, 61 randomly selected loamy ecological site plots and 39 randomly selected thin loamy ecological site plots monitored in 2012 were used for comparing selected monitoring practices.

Ecological Site Determination

Two 75m transects were laid out perpendicular to each other facing each cardinal direction at each plot location. A soil pit was dug near the center point at a location that best fitted the majority of the plot. Soil and landscape position information was used to determine the major ecological site of the plot. Plots, classified as loamy and thin loamy, were used for our analysis.

Method: Clipping by Species

A 0.178 m² hoop was placed at 20m intervals along each transect (totaling six hoops per plot). All live plant species rooted within the hoop were clipped to ground level and placed in

individual bags. Previous year's growth and litter were also collected and placed in separate bags. These bags were then oven dried at 105°C for 72 hours and weighed to the nearest 0.1 g. The average weight of each species was then determined by dividing the total weight of each species on a plot by the six clippings. This average weight was used to create a species ranking for each plot.

Method: Quadrat Frequency Frame

A 0.25 m² steel frame was placed at 10m intervals along each transect (totaling 14 frames per plot). All forb or shrub species rooted were counted within the frame at each point to determine density. A 0.1 m² frame nested within the 0.25 m² frame was used to determine presence or absence of grasses and graminoids (Curtis and McIntosh 1950, Dix 1958, Biondini et al. 1989). These presence/absence data were used to determine frequency (contacts per frame) of each grass and graminoid species, which was then used to produce a species ranking. Density of forbs and shrubs was converted to frequency to produce the same ranking.

Method: Line Point Intercept

A pin flag, approximately 1mm in diameter, was placed perpendicular to the ground at two meter intervals along the length of each transect (totaling 74 points per plot). All live plant species that touched the pin at any height were recorded, and the ground cover contacted by the point's tip was recorded. These contacts were averaged between the 74 points to produce an average frequency species ranking.

Statistical Analysis

Comparisons were made between methods within a single ecological site. Each method produced a list of species present on a plot (EXAMPLE: Table 1). This species list was sorted by frequency (line point intercept, and quadrat frame) or by average weight (clipping). This produced a

relative species ranking for each method on each plot. No data transformations were required, as the data were compared within plots. The rankings produced by line point intercept and quadrat frequency frames were compared to those produced by clipping using the Mantel Test (Mantel 1967, Monte Carlo randomization, Relative Sorenson Distance). Standard diversity measurements (species richness, evenness, and Shannon Index diversity) were also calculated and compared between methods using a student's t-test (paired, two-tailed).

Table 1. Example of relative species abundance measure produced by each method on a loamy ecological site on the Dakota Prairie Grasslands in western North Dakota in 2012.

Plot ID	Species	Common	Line point: Proportion (0-1)	Clipping (g/m ²)	Quadrat Proportion (0-1)
NDSU188ESX11-5_L	ACHMIL	Western Yarrow	0	0	0.1429
	ANTMIC	Littleleaf Pussytoes	0	0	0.1429
	ARIPUR	Purple Threawn	0	47.38	0.4286
	ARTFRI	Fringed Sagewort	0.0135	3.40	0.5
	BOUGRA	Blue Grama	0.0675	7.96	0.6429
	CALLON	Prairie Sandreed	0	6.7	0
	CARDUR	Needleleaf Sedge	0	0.71	0.1429
	CARFIL	Threadleaf Sedge	0.0541	4.45	0.7857
	HESCOM	Needle-and-Thread	0.2567	0.03	0.2857
	NASVIR	Green Needlegrass	0.0135	7.69	0.50
	PASSMI	Western Wheatgrass	0.2972	40.09	1.0
	PHLHOO	Hood's Phlox	0	0	0.2142
	POACOM	Canada Bluegrass	0	0	0.1429
	POAPRA	Kentucky Bluegrass	0	11.23	0.2857
	RATCOL	Prairie Coneflower	0.0270	0.63	0.2142
	SCHSCO	Little Bluestem	0	0	0.1429
SPHCOC	Scarlet Globemallow	0	0	0.1429	

Mantel Test Analysis

The Mantel method (Mantel 1967) “tests the significance of the correlation between matrices by evaluating results of repeated randomization” (McCune and Grace 2002). It is an alternative to regression analysis which avoids the problem of partial dependence within the matrix. This test was developed to examine clustering of diseases, but is versatile enough to be expanded to many applications, including community ecology. The Mantel Test uses a null hypothesis of no correlation between matrices (in this case species frequency rankings produced by different monitoring methods against clipping), comparing the relative distance between variables (species) rather than simply the species rankings produced. The test uses a randomization (Monte Carlo) approach to produce a standardized Mantel statistic (r) value between -1 and 1, with positive r values indicating positive correlation. A p value can be calculated from “the number of randomizations that yielded a test statistic equal to or more extreme than the observed value” (McCune and Grace 2002). We used an alpha value of 0.05 to indicate a statistically significant correlation between matrices.

RESULTS

Diversity Analysis

In order to explain and visualize possible differences between the methods, three standard diversity indices were calculated for each method on each plot (McCune and Mefford 2011). Species richness describes the total number of species found on each plot. Species evenness describes the relative population sizes of species within a plot. The Shannon index is a commonly used measure of biodiversity which incorporates both richness and evenness (Shannon 1948). These indices also provide a comparison of the depth of data provided by each monitoring method.

Loamy Ecological Site

The three methods produced different ($p \leq 0.05$) average values for richness, evenness, and Shannon Index diversity; with the exception of the line point intercept compared to clipping for Shannon Index (Table 2). On average, the quadrat frequency frame provided a greater number of plant species than the other methods.

Table 2. Standard diversity measurements calculated from each vegetation data collection method within the loamy ecological site. Shared superscript within diversity measurement type indicates no statistical difference ($n=61$, $p \leq 0.05$).

Loamy Ecological Site			
Diversity Measure	Line Point Intercept	Clipping by Species	Quadrat Frequency
Richness	11.00 ^a	13.40 ^b	18.50 ^c
Evenness	0.71 ^a	0.63 ^b	0.92 ^c
Shannon Index	1.65 ^a	1.59 ^a	2.61 ^b

Thin Loamy Ecological Site

The three methods produced different ($p \leq 0.05$) average values for richness, evenness, and Shannon Index diversity on the thin loamy ecological site, with the exception of line point intercept compared to clipping with Shannon Index (Table 3). On average, the quadrat frequency frame provided a greater number of plant species than the other methods.

Table 3. Standard diversity measurements calculated from each vegetation data collection method within the thin loamy ecological site. Shared superscript within diversity measurement type indicates no statistical difference ($n=39$, $p \leq 0.05$).

Thin Loamy Ecological Site			
Diversity Measure	Line Point Intercept	Clipping by Species	Quadrat Frequency
Richness	10.10 ^a	15.40 ^b	19.70 ^c
Evenness	0.72 ^a	0.64 ^b	0.92 ^c
Shannon Index	1.66 ^a	1.73 ^a	2.70 ^b

Loamy Ecological Site

Quadrat frequency frames and clipping by species produced a similar relative plant community ranking (H_0 : no correlation between methods, $r: 0.609$, $p=0.001$) for the loamy ecological site. Line point intercept and clipping by species also produced a similar relative plant community ranking ($r = 0.769$, $p=0.001$). Nonmetric multidimensional scaling (NMS) was used to produce a visual representation of the Mantel Test results (Figure 1). The directional vectors in the graph connect the rankings produced by each method within a plot.

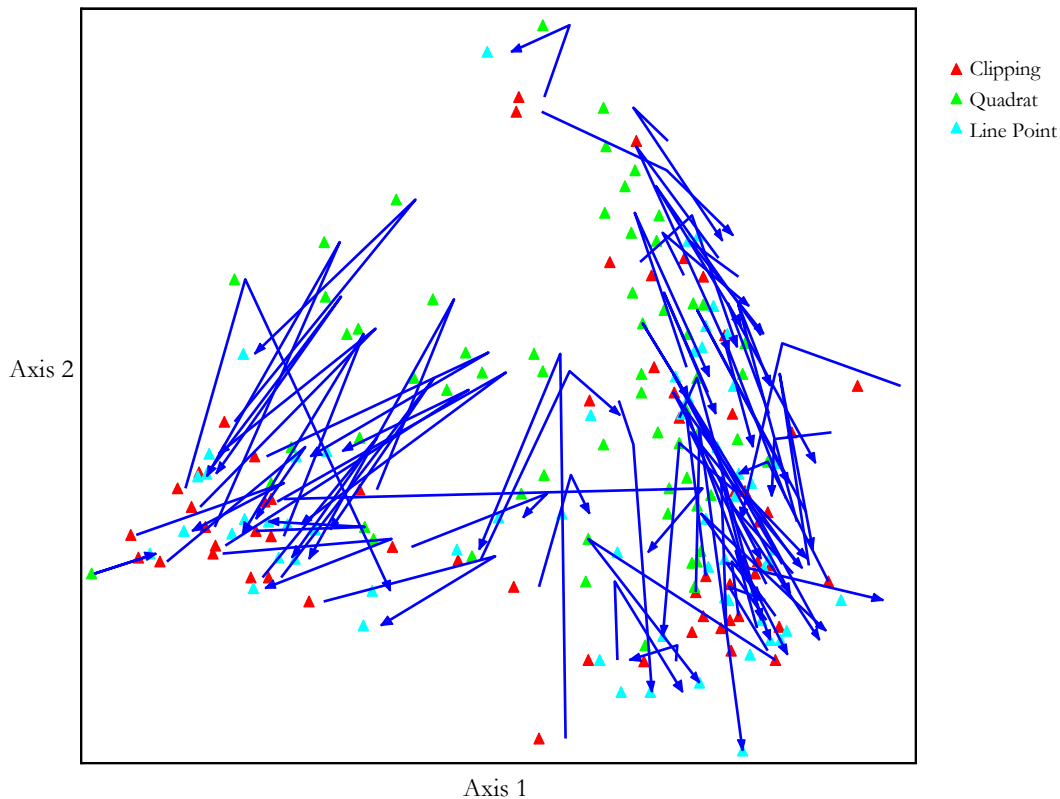


Figure 3. Nonmetric multidimensional scaling plot with directional vectors connecting results of the three vegetation data collection methods within a plot on the loamy ecological site in western North and South Dakota in 2012.

Thin Loamy Ecological Site

Quadrat frequency frames and clipping by species produced a similar relative plant community ranking (H_0 : no correlation between methods, $r = 0.719$, $p=0.001$) for the thin loamy ecological site. Line point intercept and clipping by species also produced a similar relative plant community ranking ($r: 0.819$, $p=0.001$). Nonmetric multidimensional scaling was used to produce a visual representation of the Mantel Test results (Figure 2). The directional vectors in the graph connect the rankings produced by each method within a plot.

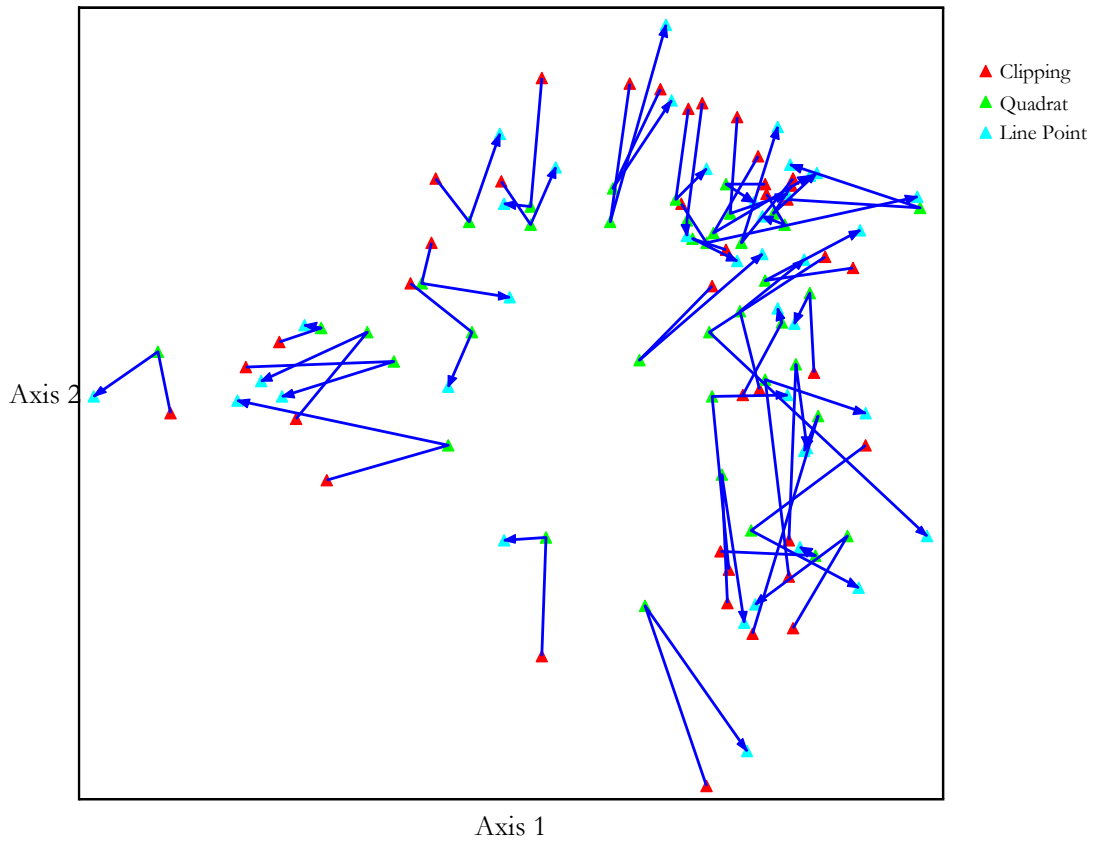


Figure 4. Nonmetric multidimensional scaling plot with directional vectors connecting results of the three vegetation data collection methods (1=Clipping, 2=Quadrat Frame, 3=Line Point Intercept) within a plot on the thin loamy ecological site in western North and South Dakota in 2012.

DISCUSSION

Diversity Analysis

The diversity analyses indicate that each method produces a statistically different result for commonly used diversity measurements. This is important if researchers or managers are using an indicator such as richness or diversity to make management decisions. In a case like this, whichever data collection method was used in a project, the same method should be used for all monitoring in order to avoid method bias.

Another important finding was the ability of the quadrat frequency frame to detect more species (i.e., higher richness values). This may seem intuitive due to the larger area covered by the frequency frame method when compared to clipping; however, clipping is still often chosen as a default monitoring method despite this disadvantage. Based on our findings, if detecting rare species or producing a complete species presence list is the goal, the quadrat frequency frame technique is the best method of the three. Stohlgren et al. (1998) also found that each of his methods (Daubenmire, Modified-Whittaker, large quadrat, and Parker) produced a different value for species richness, with the large quadrat (1.0 m²) being the most cost efficient method.

Previous Comparisons

Few detailed comparisons of these methods have been done before. Prosser et al. (2003) compared a 0.25 m² quadrat frame to the U.S. Army Land Condition-Trend Analysis (LCTA) Program, which utilizes Line Point Intercept as a core method. They found that quadrat frequency frames better portrayed community diversity and species composition in a complex mixed grass prairie ecosystem. This finding agrees with our results when species richness is considered. Walker (1970) found that quadrat frequency frames were a simple to use procedure that provided useful

data without an impractical time input, but West (1985) warned that managers should not attempt to calculate percent composition from frequency data, and Hironaka (1985) stated that other methods should be used in conjunction with quadrat frequency if time and money are available. Clark et al. (2008) compared the point intercept method with clipping in sagebrush steppe ecosystems and found point intercept as accurate as clipping but less time consuming, which agrees with our analysis.

Mantel Analysis

The findings of our Mantel test analysis show a strong correlation between the relative species rankings produced by the three monitoring methods. In all comparisons, this correlation was significant. This indicates that on loamy and thin loamy ecological sites in the northern Mixed-Grass Prairie, either of the less time-consuming methods (line point intercept and quadrat frequency frames), will produce the same relative species ranking as clipping by species.

When working on a landscape level, relative species rankings should prove adequate for making management decisions. These lists show which are dominant species, which are rare species, give an estimate of species richness, and provide information on relative proportions of desirable and undesirable species. There were differences in diversity index values produced by the three methods, but we believe that our similarity based analysis provides a more valuable tool for comparing rangeland sites. While the USDA-NRCS advocates using actual species weight (clipping by species) as a method for using similarity indices and state-transitional models (USDA-NRCS 1997), our results indicate that quadrat frequency frames and line point intercept have the potential to fit well with these models.

CONCLUSION

Relative species ranking, generated by quadrat frequency frames, is an efficient, cost-effective, and accurate way of monitoring mixed-grass prairie. We recommend that researchers and managers use careful consideration when choosing between rangeland monitoring methods. Too often we choose the most intensive method on the premise that it is also the best. We found that quadrat frequency frames and line point intercept produced the same ranking of species abundance as clipping, and did so with less time input. Quadrat frequency frames also produced a more complete survey of the species present on each plot. The quadrat frequency method has been criticized but we feel that the benefits of the method outweigh the possible limitations in many cases, especially when coupled with relative species ranking. Researchers and managers can often save time and money using these methods without sacrificing critical data quality. Our data also suggest that species abundance rankings could be used to compare previously surveyed sites. This has some promise for researchers needing to compare species composition on sites that were surveyed using different methodologies. This method of comparison needs further study to validate its use in different areas and ecosystems.

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APPENDIX A: PLANT SPECIES OCCURRING ON LOAMY ECOLOGICAL SITE

PLOTS IN 2012 (130 SPP.)

<i>Achillea millefolium</i>	Western Yarrow	<i>Cirsium undulata</i>	Wavyleaf Thistle
<i>Agropyron cristatum</i>	Crested Wheatgrass	<i>Comata umbellata</i>	Bastard Toadflax
<i>Allium textile</i>	Textile Onion	<i>Convovulvis arvensis</i>	Field Bindweed
<i>Andropogon gerardii</i>	Big Bluestem	<i>Coryphantha vivipara</i>	Spinystar
<i>Anemone cylindrica</i>	Candle Anemone	<i>Crepis acuminata</i>	Tapertip Hawksbeard
<i>Antennaria</i>		<i>Dalea purpurea</i>	Purple Prairie Clover
<i>microcephala</i>	Littleleaf Pussytoes	<i>Dasiphora fruticosa</i>	Shrubby Cinquefoil
<i>Antennaria neglecta</i>	Field Pussytoes	<i>Descurainia sophioides</i>	Northern Tansymustard
<i>Antennaria parvifolia</i>	Small-Leaf Pussytoes	<i>Dichanthelium</i>	
<i>Aristida longespica</i>	Slimspike Threawn	<i>oligosanthes</i>	Scribner's Rosette Grass
<i>Aristida purpurea</i>	Purple Threawn	<i>Dichanthelium</i>	
<i>Arnica fulgens</i>	Foothill Arnica	<i>wilcoxianum</i>	Fall Rosettegrass
<i>Artemisia canadensis</i>	Silver Sagebrush	<i>Distichlis spicat</i>	Inland Saltgrass
<i>Artemisia</i>		<i>Dodecatheon</i>	
<i>dracunculoides</i>	Green Sagewort	<i>pulchellum</i>	Darkthroat Shootingstar
<i>Artemisia frigida</i>	Fringed Sagewort	<i>Echinacea angustifolia</i>	Black Samson
<i>Artemisia ludoviciana</i>	White Sagewort	<i>Elymus lanceolatus</i>	Thickspike Wheatgrass
<i>Artemisia tridentata</i>	Big Sagebrush	<i>Elymus trachycaulus</i>	Slender Wheatgrass
<i>Asclepias pumilla</i>	Plains Milkweed	<i>Erigeron bellidiastrum</i>	Daisy Fleabane
<i>Astragalus agrestis</i>	Purple Milkvetch	<i>Eriogonum flavum</i>	Alpine Golden Buckwheat
<i>Astragalus laxmannii</i>	Prairie Milkvetch	<i>Erigeron strigosus</i>	Prairie Fleabane
<i>Astragalus missouriensis</i>	Missouri Milkvetch	<i>Escobaria vivipara</i>	Spinystar
<i>Atriplex nuttalli</i>	Nuttall's Saltbush	<i>Euphorbia esula</i>	Leafy Spurge
<i>Bouteloua curtipendula</i>	Sideoats Grama	<i>Galium boreale</i>	Northern Bedstraw
<i>Bouteloua gracilis</i>	Blue Grama	<i>Gaura coccinea</i>	Scarlet Beeblossom
<i>Bromus arvensis</i>	Field Brome	<i>Geum triflorum</i>	Prairie Smoke
<i>Bromus inermis</i>	Smooth Brome	<i>Grindelia squarrosa</i>	Curlycup Gumweed
<i>Bromus tectorum</i>	Cheatgrass	<i>Gutierrezia sarothrae</i>	Broom Snakeweed
<i>Buchloe dactyloides</i>	Buffalograss	<i>Hedeoma hispida</i>	False Pennyroyal
<i>Calamovilfa longifolia</i>	Prairie Sandreed	<i>Helianthus pauciflora</i>	Stiff Sunflower
<i>Calamagrostis</i>		<i>Hesperostipa comata</i>	Needle-and-Thread
<i>montanensis</i>	Plains Reedgrass	<i>Hesperostipa spartina</i>	Porcupinegrass
<i>Campanula rotundifolia</i>	Bluebell Bellflower	<i>Heterotheca villosa</i>	Hairy False Goldenaster
<i>Carex duriuscula</i>	Needleleaf Sedge	<i>Juncus horizontalis</i>	Creeping Juniper
<i>Carex filifolia</i>	Threadleaf Sedge	<i>Koeleria macrantha</i>	Prairie Junegrass
<i>Carex inops</i>	Sun Sedge	<i>Krascheninnikovia</i>	
<i>Cerastium arvense</i>	Field Chickweed	<i>lanata</i>	Winterfat
<i>Cirsium arvense</i>	Canada Thistle	<i>Lactuca tatarica</i>	Blue Lettuce
<i>Cirsium flodmanii</i>	Flodman's Thistle	<i>Lesquerella arenosa</i>	Great Plains Bladderpod

<i>Liatriis punctata</i>	Dotted Blazingstar	<i>Pulsatilla patens</i>	Pasquflower
<i>Linus lewisii</i>	Prairie Flax	<i>Punctelia species</i>	Punctelia
<i>Lithospermum incisum</i>	Narrowleaf Stoneseed	<i>Ratibita columnifera</i>	Prairie Coneflower
<i>Lygodesmia juncea</i>	Rush Skeletonplant	<i>Rhus trilobata</i>	Skunkbush Sumac
<i>Machaeranthera pinnatifida</i>	Lacy Tansyaster	<i>Rosa arkana</i>	Prairie Rose
<i>Medicago lupulina</i>	Black Medick	<i>Schedonnardus paniculatus</i>	Tumblegrass
<i>Medicago sativa</i>	Alfalfa	<i>Schizachirium scoparium</i>	Little Bluestem
<i>Melilotus officinalis</i>	Yellow Sweetclover	<i>Selaginella densa</i>	Clubmoss
<i>Monarda fistulosa</i>	Wild Bergamot	<i>Shepherdia argentea</i>	Silver Buffaloberry
<i>Muhlenbergia cuspidata</i>	Plains Muhly	<i>Sisyrinchium montanum</i>	Strict Blue-eyed Grass
<i>Nassella viridula</i>	Green Needlegrass	<i>Solidago missouriensis</i>	Missouri Goldenrod
<i>Oligoneuron album</i>	Prairie Goldenrod	<i>Solidago mollismus</i>	Velvety Goldenrod
<i>Oligoneuron rigidum</i>	Stiff Goldenrod	<i>Spharalcea coccinea</i>	Scarlet Globemallow
<i>Opuntia fragilis</i>	Brittle Pricklypear	<i>Sporobolus cryptandrus</i>	Sand Dropseed
<i>Opuntia polyacantha</i>	Plains Pricklypear	<i>Sporobolus heterolepis</i>	Prairie Dropseed
<i>Oxytropis lambertii</i>	Lambert's Crazyweed	<i>Symphyotrichum ericoides</i>	White Heath Aster
<i>Pascopyrum smithii</i>	Western Wheatgrass	<i>Symphyotrichum falcatum</i>	White Prairie Aster
<i>Pediomelum aromaticum</i>	Silverleaf Scurfpea	<i>Symphyotrichum laeve</i>	Smooth Blue Aster
<i>Pediomelum esculentum</i>	Indian Breadroot	<i>Symphyotrichum oblongifolia</i>	Aromatic Aster
<i>Phlox hoodii</i>	Hood's Phlox	<i>Symphoricarpos occidentalis</i>	Western Snowberry
<i>Plantago patagonia</i>	Wooly Plantain	<i>Taraxacum officinale</i>	Dandelion
<i>Poa compressa</i>	Canada Bluegrass	<i>Tetranneuris acaulis</i>	Stemless Four-Nerve Daisy
<i>Poa pratensis</i>	Kentucky Bluegrass	<i>Thermopsis gracilis</i>	Slender Goldenbanner
<i>Poa secunda</i>	Sandberg Bluegrass	<i>Thinopyrum intermedium</i>	Intermediate Wheatgrass
<i>Polygala alba</i>	White Milkwort	<i>Tragopogon dubius</i>	Goatsbeard
<i>Polygonum douglasii</i>	Douglas' Knotweed	<i>Vicia americana</i>	American Vetch
<i>Portulaca species</i>	Purslane Species	<i>Viola nuttallii</i>	Nuttall's Violet
<i>Potentilla effusa</i>	Branched Cinquefoil	<i>Vulpia octoflora</i>	Sixweeks Fescue
<i>Potentilla pensylvanica</i>	Pennsylvania Cinquefoil		
<i>Potentilla pulchella</i>	Pretty Cinquefoil		
<i>Prunus virginianus</i>	Chokecherry		

APPENDIX B: PLANT SPECIES OCCURRING ON THIN LOAMY ECOLOGICAL

SITE PLOTS IN 2012 (143 SPP.)

<i>Achillea millefolium</i>	Western Yarrow	<i>Campanula rotundifolia</i>	Bluebell Bellflower
<i>Agropyron cristatum</i>	Crested Wheatgrass	<i>Carex duriuscula</i>	Needleleaf Sedge
<i>Allium textile</i>	White Wild Onion	<i>Carex filifolia</i>	Threadleaf Sedge
<i>Amorpha canadensis</i>	Leadplant	<i>Carex inops</i>	Sun Sedge
<i>Andropogon gerardii</i>	Big Bluestem	<i>Carex sprengelii</i>	Sprengel's Sedge
<i>Anemone canadensis</i>	Canadian Anemone	<i>Cerastium arvense</i>	Field Chickweed
<i>Anemone cylindrica</i>	Candle Anemone	<i>Cirsium flodmanii</i>	Flodman's Thistle
<i>Antennaria microcephala</i>	Littleleaf Pussytoes	<i>Cirsium undulata</i>	Wavyleaf Thistle
<i>Antennaria neglecta</i>	Field Pussytoes	<i>Comata umbellata</i>	Bastard Toadflax
<i>Antennaria parvifolia</i>	Small-Leaf Pussytoes	<i>Convolvulus arvensis</i>	Field Bindweed
<i>Aristida purpurea</i>	Purple Threeawn	<i>Conyza canadensis</i>	Canadian Horseweed
<i>Arnica fulgens</i>	Foothill Arnica	<i>Dalea purpurea</i>	Purple Prairie Clover
<i>Artemisia canadensis</i>	Silver Sagebrush	<i>Dasiphora fruticosa</i>	Shrubby Cinquefoil
<i>Artemisia dracunculoides</i>	Green Sagewort	<i>Dichanthelium leibergii</i>	Leiberg's Panicum
<i>Artemisia frigida</i>	Fringed Sagewort	<i>Dichanthelium oligosanthes</i>	Scribner's Rosette Grass
<i>Artemisia ludoviciana</i>	White Sagewort	<i>Distichlis spicata</i>	Inland Saltgrass
<i>Artemisia tridentata</i>	Big Sagebrush	<i>Dodecatheon pulchellum</i>	Darkthroat Shootingstar
<i>Asclepias pumilla</i>	Plains Milkweed	<i>Echinacea angustifolia</i>	Black Samson
<i>Asclepias verticillata</i>	Whorled Milkweed	<i>Elymus lanceolatus</i>	Thickspike Wheatgrass
<i>Astragalus agrestis</i>	Purple Milkvetch	<i>Elymus repens</i>	Quackgrass
<i>Astragalus americanus</i>	American Milkvetch	<i>Elymus trachycaulus</i>	Slender Wheatgrass
<i>Astragalus gilviflorus</i>	Plains Milkvetch	<i>Eriogonum flavum</i>	Alpine Golden Buckwheat
<i>Astragalus laxmannii</i>	Prairie Milkvetch	<i>Erigeron strigosus</i>	Prairie Fleabane
<i>Astragalus missouriensis</i>	Missouri Milkvetch	<i>Escobaria vivipara</i>	Spinystar
<i>Atriplex nuttalli</i>	Nuttall's Saltbush	<i>Euphorbia esula</i>	Leafy Spurge
<i>Balsamorhiza sagittata</i>	Arrowleaf Balsamroot	<i>Galium boreale</i>	Northern Bedstraw
<i>Bouteloua curtipendula</i>	Sideoats Grama	<i>Gaura coccinea</i>	Scarlet Beeblossom
<i>Bouteloua gracilis</i>	Blue Grama	<i>Geum triflorum</i>	Prairie Smoke
<i>Bouteloua hirsuta</i>	Hairy Grama	<i>Glycyrrhiza lepidota</i>	American Licorice
<i>Bromus arvensis</i>	Field Brome	<i>Grindelia squarrosa</i>	Curlycup Gumweed
<i>Bromus inermis</i>	Smooth Brome	<i>Gutierrezia sarothrae</i>	Broom Snakeweed
<i>Bromus tectorum</i>	Cheatgrass	<i>Hedeoma hispida</i>	False Pennyroyal
<i>Buchloe dactyloides</i>	Buffalograss	<i>Helianthus pauciflora</i>	Stiff Sunflower
<i>Buglossoides arvensis</i>	Corn Gromwell	<i>Hesperostipa comata</i>	Needle-and-Thread
<i>Calamovilfa longifolia</i>	Prairie Sandreed	<i>Hesperostipa spartina</i>	Porcupinegrass
<i>Calamagrostis montanensis</i>	Plains Reedgrass	<i>Heterotheca villosa</i>	Hairy False Goldenaster
<i>Calystegia sericata</i>	False Bindweed	<i>Juniperus communis</i>	Common Juniper

<i>Juniperus horizontalis</i>	Creeping Juniper	<i>Potentilla pensylvanica</i>	Pennsylvania Cinquefoil
<i>Juniperus scopulorum</i>	Rocky Mountain Juniper	<i>Prunus virginiana</i>	Chokecherry
<i>Koeleria macrantha</i>	Prairie Junegrass	<i>Pulsatilla patens</i>	Cutleaf Anemone
<i>Krascheninnikovia lanata</i>	Winterfat	<i>Ratibida columnifera</i>	Prairie Coneflower
<i>Lactuca tatarica</i>	Blue Lettuce	<i>Rhus trilobata</i>	Skunkbrush Sumac
<i>Liatris punctata</i>	Dotted Blazingstar	<i>Rosa arkana</i>	Prairie Rose
<i>Linus lewisii</i>	Prairie Flax	<i>Schizachirium scoparium</i>	Little Bluestem
<i>Lithospermum canescens</i>	Hoary Puccoon	<i>Selaginella densa</i>	Clubmoss
<i>Lithospermum incisum</i>	Narrowleaf Stoneseed	<i>Senecio spp.</i>	Ragwort
<i>Lomatium foeniculaceum</i>	Biscuitroot	<i>Shepherdia argentea</i>	Silver Buffaloberry
<i>Lygodesmia juncea</i>	Rush Skeletonweed	<i>Sisyrinchium montanum</i>	Strict Blue-eyed Grass
<i>Medicago sativa</i>	Alfalfa	<i>Solidago missouriensis</i>	Missouri Goldenrod
<i>Melilotus officinalis</i>	Yellow Sweetclover	<i>Solidago mollismus</i>	Velvety Goldenrod
<i>Monarda fistulosa</i>	Wild Bergamot	<i>Solidago nemoralis</i>	Gray Goldenrod
<i>Muhlenbergia cuspidata</i>	Plains Muhly	<i>Solidago speciosa</i>	Showy Goldenrod
<i>Nassella viridula</i>	Green Needlegrass	<i>Spartina pectinata</i>	Prairie Cordgrass
<i>Oenothera flava</i>	Yellow Evening Primrose	<i>Spharalcea coccinea</i>	Scarlet Globemallow
<i>Oligoneuron album</i>	Prairie Goldenrod	<i>Sporobolus cryptandrus</i>	Sand Dropseed
<i>Oligoneuron rigidum</i>	Stiff Goldenrod	<i>Sporobolus heterolepis</i>	Prairie Dropseed
<i>Opuntia fragilis</i>	Brittle Pricklypear	<i>Symphyotrichum ciliolatum</i>	Alkali Aster
<i>Opuntia polyacantha</i>	Plains Pricklypear	<i>Symphyotrichum ericoides</i>	White Heath Aster
<i>Oxytropis lambertii</i>	Purple Locoweed	<i>Symphyotrichum falcatum</i>	White Prairie Aster
<i>Pascopyrum smithii</i>	Western Wheatgrass	<i>Symphyotrichum laeve</i>	Smooth Blue Aster
<i>Pediomelum argophyllum</i>	Silverleaf Scurfpea	<i>Symphyotrichum oblongifolia</i>	Aromatic Aster
<i>Pediomelum esculentum</i>	Indian Breadroot	<i>Symphoricarpos occidentalis</i>	Western Snowberry
<i>Penstemon angustifolius</i>	Beardtongue	<i>Taraxacum officinale</i>	Dandelion
<i>Penstemon glaber</i>	Sawsepal Penstemon	<i>Tetraneuris acaulis</i>	Stemless Four-Nerve Daisy
<i>Phlox hoodii</i>	Hood's Phlox	<i>Thinopyrum intermedium</i>	Intermediate Wheatgrass
<i>Plantago patagonia</i>	Wooly Plantain	<i>Toxicodendron rydbergii</i>	Poison Ivy
<i>Poa compressa</i>	Canada Bluegrass	<i>Tragopogon dubius</i>	Goatsbeard
<i>Poa pratensis</i>	Kentucky Bluegrass	<i>Vicia americana</i>	American Vetch
<i>Poa secunda</i>	Sandberg Bluegrass	<i>Viola pedatifida</i>	Prairie Violet
<i>Polygala alba</i>	White Milkwort	<i>Yucca glauca</i>	Soapweed Yucca
<i>Potentilla effusa</i>	Branched Cinquefoil	<i>Zizia aptera</i>	Meadow Zizia
<i>Potentilla hippiana</i>	Wooly Cinquefoil		

APPENDIX C: LITTLE MISSOURI NATIONAL GRASSLAND ALLOTMENTS USED

FOR ANALYSIS FROM DATA COLLECTED IN 2012

Loamy Ecological Site	
Allotment	Plots Used
1	1
13	1
37	2
43	2
51	3
53	3
54	2
58	1
69	1
84	2
108	2
120	1
165	2
166	1
175	1
186	2
188	2
190	1
191	1
202	2
273	5
275	1
370	1
371	3
374	2
376	2
405	1
407	1
408	2
409	1
411	2
412	2
512	5

Thin Loamy Ecological Site	
Allotment	Plots Used
1	1
27	1
33	1
37	2
53	2
54	3
69	2
108	3
109	1
115	1
165	1
166	5
169	5
186	2
188	2
190	2
202	1
273	4